

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

1

2

3

4

5

6

7

8

9

10

11

12

13

14

15

16

17

18

19

20

21

22

23

24

25

26

27

28

29

30

31

32

33

34

35

36

37

38

39

40

41

42

43

44

45

46

47

48

49

50

51

52

53

54

55

56

57

58

59

60

Food availability and population structure: How do clumped and abundant sources of carrion affect the genetic diversity of the black-backed jackal?

Robert S. James*. School of Pharmacy and Bioscience, University of Brighton, Brighton, BN2 4GJ, UK.
R.S.James@brighton.ac.uk

Dawn M. Scott. School of Pharmacy and Bioscience, University of Brighton, Brighton, BN2 4GJ, UK.
Dawn.scott@brighton.ac.uk

Richard W. Yarnell. School of Animal Rural & Environmental Sciences, Nottingham Trent University, Southwell, NG25 0QF
Richard.Yarnell@NTU.ac.uk

Andrew D. J. Overall. School of Pharmacy and Bioscience, University of Brighton, BN2 4GJ, UK.
A.D.J.Overall@brighton.ac.uk

* Corresponding author.

Key Words:

Black-backed jackal, supplementary feeding, scavenger, non-invasive genetic sampling, microsatellite, molecular ecology, population genetics, zoology.

Abstract

Carnivores frequently come into conflict with humans in agricultural and livestock producing areas around the world. Understanding their fidelity and dispersal patterns in response to food availability is therefore important given the effort invested in conflict mitigation strategies. In this study, we investigated the influence of clumped and abundant sources of carrion on the genetic diversity of the black-backed jackal (*Canis mesomelas*) within six private game farms in the North West and Gauteng provinces of South Africa. It is predicted that clumped and abundant sources of carrion will increase immigration and thus genetic diversity in the local subpopulation. By quantifying the variability of microsatellite loci in black-backed jackals subjected to artificially increased carrion availability, and comparing them with individuals from control sites, we were able to describe patterns of historic gene flow within the total sampled population. The results of this investigation indicate that clumped and abundant sources of carrion promote genetic structuring ($F_{ST} = 0.0302$) which implies a lack of gene flow and a degree of isolation. Genetic artefacts of three populations could be identified through Bayesian clustering analysis of individuals based on their genetic identity. Individuals sampled from the two supplementary feeding sites could be assigned to one of two ancestral populations with an average population assignment of 69% and 82%, while individuals from the remaining four control sites, originate from a third population with percentage assignments of 63%, 46%, 53% and 42%. It is therefore likely that clumped and abundant sources of carrion in the agricultural landscape of South Africa can affect the population dynamics of the black-backed jackal and result in subpopulations with limited migration and dispersal when compared with the total population.

Introduction

It is generally recognised that carnivores play a fundamental role in the structure and function of an ecosystem (Ripple, et al., 2014; Ripple & Beschta, 2004). However, factors such as disease transmission and livestock depredation frequently promote conflict in areas where humans and carnivores exist in close proximity (Woodroffe, et al., 2005). Understanding the ecological factors that drive the spatial organisation of free-ranging carnivores is therefore important when considering both conservation and management of species in the human-modified landscape. Thus this study follows a microsatellite-based approach to investigate the short term historic effects of four years of supplementary feeding on the genetic diversity of black-backed jackals (*Canis mesomelas*) at private game farms in South Africa.

Following the expectations of the resource dispersion hypothesis (Macdonald, 1983), an increase in localised food availability will often result in a breakdown in territorial stability and subsequently lead to an increase in local density (Johnson, et al., 2002; Johnson, et al., 2001). Indeed, anthropogenically derived sources of food, synonymous with agricultural and human modified landscapes, have been shown to strongly influence the spatial organisation of many omnivorous canids including the golden jackal (*Canis aureus* Rotem, et al., 2011), red fox (*Vulpes vulpes* Contesse, et al., 2004) coyote (*Canis latrans* Fedriani, et al., 2001) and dingo (*Canis lupus dingo* Newsome, et al., 2013). Furthermore, studies in both Namibia and South Africa have recorded the black-backed jackal at far greater abundances than expected in areas where scavenging opportunities are high and carrion availability is clumped, stable

1
2
3 74 and abundant (Yarnell, et al., 2014; Jenner, et al., 2001; Hiscocks & Perrin, 1988). Studies
4
5 75 using both radio-telemetry and behavioural observations in the Cape Cross Seal Reserve
6
7 76 (CCSR) have also concluded that territorial boundaries of the black-backed jackal often
8
9 77 overlap in close proximity to clumped, abundant resources such as seal colonies (Hiscocks &
10
11 78 Perrin, 1988), and that home range sizes significantly increase with distance from the colony
12
13 79 itself (Jenner, et al., 2001). As the social structure of the black-backed jackal is commonly
14
15 80 reported to consist of a monogamous breeding pair, which holds and aggressively defends
16
17 81 territory from transient individuals and neighbouring residents (Estes, 1991; Ferguson, et al.,
18
19 82 1983), it is clear that an increase in local abundance of food can dramatically affect both the
20
21 83 territorial behaviour and spatial organisation of this species. However, what remains unclear
22
23 84 from contemporary observations is the effect that increased food availability has on the
24
25 85 fidelity and dispersal of such subpopulations over time. Therefore by examining the genetic
26
27 86 diversity of black-backed jackals in the game farms of South Africa, this study aims to
28
29 87 elucidate the genetic consequences of clumped and abundant sources of food on the
30
31 88 dispersal of a free-ranging canid within a human-modified landscape.
32
33
34
35
36
37
38
39
40
41

42 90 The black-backed jackal is a medium sized canid (5-15 kg) with two discrete distributions
43
44 91 that span the majority of the Southern African sub-region, and parts of Eastern Africa
45
46 92 (Skinner & Chimimba, 2005; Estes, 1991). This study focuses on the southern African
47
48 93 subspecies (*C. m. mesomelas*), henceforth “black-backed jackal”, due to the high rate of
49
50 94 human-carnivore conflict associated with this region (Thorn, et al., 2012). As a vector of
51
52 1
53
54 95 rabies and canine distemper (Bellan, et al., 2012; Zulu, et al., 2009), and an opportunistic
55
56 96 hunter of small game and livestock (Estes, 1991), the black-backed jackal is frequently
57
58
59
60

2
3 97
4
5 98 controlled as a pest species throughout its range (Thorn, et al., 2012; Ginsberg &
6
7
8 99 Macdonald, 2004). With an omnivorous diet consisting of small mammals, livestock, forage
9
10 100 and carrion (Klare, et al., 2010), this species is considered a generalist carnivore that is able
11
12 to undertake diet switching in response to changes in local food availability (Kamler, et al.,
13 101
14 2012; van der Merwe, et al., 2009; Rowe-Rowe, 1983; Fourie, et al., 2015; Humphries, et al.,
15 102
16 2016). Therefore, to further investigate the effect of food availability on the population
17 103
18 dynamics of the black-backed jackal, this study used carrion stations, known as vulture
19 104
20 restaurants, to measure the historic effect of artificially increasing scavenging material on
21 105
22 the gene flow and variation in genetic diversity within and between local subpopulations.
23 106
24 Vulture restaurants were originally introduced in participating game farms and nature
25 107
26 reserves across South Africa with an aim to supply declining vulture species with a safe and
27 108
28 consistent source of carrion which originates from hunted or slaughtered livestock destined
29 109
30 for the human food chain. Subsequent analysis has shown that the regular deposition of
31 110
32 carcasses at these sites has resulted in an unintentional increase in the local abundance of
33 111
34 many scavenging carnivores, including the black-backed jackal (Yarnell, et al., 2014). As the
35
36 abundance of black-backed jackals residing in close proximity to vulture feeding sites are
37 112
38 often far in excess of those in the surrounding area (pers. obs.), it is predicted that clumped
39 113
40 and abundant sources of carrion will have resulted in an increase in genetic diversity within
41 114
42 local subpopulations as it is hypothesised that increased food availability increases
43 115
44
45
46
47 116 migration.

48
49
50
51 117

52 53 54 118 **Methods**

55
56
57
58 119 *Sampling and study sites*
59
60

1
2
3 120 This study was undertaken in the North-West and Gauteng provinces of South Africa.
4
5 121 Individual black-backed jackals (n = 65) were sampled for genetic material from six game
6
7 122 breeding farms (Fig. 1) between March 2011 and September 2012 for an analysis of
8
9 123 population structure. Two game farms, Site VR1 and Site VR2, had active vulture restaurants
10
11 124 initiated approximately four years prior to sampling (n = 27 and 19 jackal DNA samples,
12
13 125 respectively). The remaining four game farms, Site C1, C2, C3 and C4, acted as control sites
14
15 126 with no additional scavenging material provided (n = 6, 6, 3 and 4). Carrion, consisting of
16
17 127 recently deceased ungulates, was placed at each vulture restaurant on a regular basis with
18
19 128 an average of 797 kg a month being recorded between 2008 and 2011 at site VR1 (Yarnell,
20
21 129 et al., 2013). A non-invasive genetic recovery protocol was used to acquire genetic material
22
23 130 from 63 recently deposited faecal samples along with two tissue biopsies opportunistically
24
25 131 collected from the ear lobe of deceased individuals. The non-invasive genetic recovery
26
27 132 protocol used in this investigation was specifically designed for use with this species and had
28
29 133 previously been tested for adequate recovery of host DNA prior to undertaking analysis
30
31 134 (James, et al., 2015). Tissue samples were placed in 1.5 ml of absolute ethanol (EtOH) after
32
33 135 collection and stored at -20°C prior to transport to the UK for further analysis.
34
35
36
37
38
39
40
41

42 136
43
44
45 137 *Figure 1 approximately here.*
46

47
48 138
49
50
51 139 To sample faecal deposits for genetic source material, driven transects of 5 km were
52
53 140 undertaken along the road networks within each site. Transect routes were chosen to
54
55 141 maximise an even coverage of area and habitat types. Transect width was standardised at
56
57 142 2m from the edge of the road to minimise the variation in detection probability. All
58
59
60

1
2
3 143 transects were undertaken by two experienced observers and were driven at a speed
4
5 144 maintained between 5 and 10 km/h to maintain sampling effort. Sampling effort was
6
7
8 145 maintained between sites at 1.4 km of transect driven per 1 km² of reserve area. Upon
9
10 146 discovery of fresh faecal material, the outer most layers of the faecal sample were collected
11
12
13 147 using a sterile razor blade and stored in a biologically inert buffer (Roche diagnostics S.T.A.R.
14
15 148 buffer cat no: 03335208001). Samples were then stored at – 20°C prior to DNA extraction
16
17 149 and purification. Scat identification was aided with field guides and expert advice where
18
19 150 necessary, and the spatial location of each faecal sample was recorded using a Garmin
20
21 151 GPSmap 62 (supplementary material).
22
23
24
25
26
27
28

29 *Microsatellite loci*

30
31 154 Previous research has successfully used domestic dog (*Canis lupus familiaris*) microsatellite
32
33 155 markers to describe the genetic structure and dispersal of jackal populations (e.g. Jenner,
34
35 156 2007; Minnie 2016). However, the markers used for the current study were specifically
36
37 157 characterised for the black-backed jackal (Table 1; James, et al., 2015) and examined for
38
39 158 selective neutrality before estimates of population structure were undertaken.
40
41
42 159 Furthermore, the predictive power, resolution and allelic drop-out rate and null allele
43
44 160 estimates were evaluated for this marker set and were shown to be suitable for use in this
45
46
47 161 analysis. These markers were used to estimate the population structure and inbreeding
48
49 162 coefficients of the black-backed jackal. Individual multilocus genotype profiles that matched
50
51 163 were considered to derive from the same source and were hence removed prior to the
52
53
54 164 analysis. Results were pooled by site for an analysis of population structure.
55
56
57
58
59
60

1
2
3 **Table 1.** Microsatellite loci, 5' modification, forward (F) and reverse (R) primer sequences (5'-
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
165
166
3'), T_m and NCBI accession numbers (AN).

Locus	5' mod	F primer	T_m °C	R Primer	T_m °C	AN
cme144	FAM	aactttaagccacacttctgca	57.9	acttgcctctggcttttaagc	58.4	KU050829
cme136	FAM	aactggccaacataaacacg	58.5	ttcattaaccctttgcctg	58.5	KU050830
cme206	HEX	cgagagcaacataggcatga	58.4	caaagtgctgtggcaggtc	58.8	KU050831
cme196	HEX	aggaggacagaaagacagaagg	57.5	atggatgtattgtagggtgg	58.0	KU050832
cme193	FAM	gagctcctgatggaagagctta	58.6	catcctgtccgtgacttcaa	58.0	KU050833
cme210	HEX	cttgtgcaatcatcatcttga	57.2	cccaggtacctatggct	57.5	KU050834

167

168 *Sample processing and PCR conditions*

169 Approximately 25 mg of black-backed jackal tissue, fixed in absolute EtOH, underwent DNA
170 extraction using the Qiagen DNeasy blood and Tissue Kit (cat No: 69504), following the
171 manufacturer-based spin column tissue extraction protocol. Dermal and epidermal cells
172 were isolated manually from cartilaginous tissue before proteinase K digestion at 56°C. DNA
173 was then eluted using 150 µl of manufacturer-supplied PCR-compatible buffer.

174

175 A chloroform extraction protocol was used in conjunction with a Qiagen DNeasy spin
176 column method to isolate and purify DNA templates from faecal samples collected in the
177 field. Faecal samples stored in S.T.A.R. buffer were defrosted in batches at 4°C prior to DNA
178 extraction. Individual samples were homogenised by shaking, then 10 ml of sample was
179 transferred to a sterile collection tube. One millilitre of ≥ 99.8 % chloroform-EtOH (GC) was

1
2
3 180 then mixed with the sample solution and vortexed to form an emulsion. Emulsified samples
4
5 181 then underwent centrifugation at 1,000 x g for 3 min and the subsequent supernatant was
6
7
8 182 removed for further processing. A Qiagen blood and tissue extraction protocol was followed
9
10 183 to recover DNA from the supernatant. Spin columns were centrifuged at 1400 x g for 3 min
11
12 184 prior to elution, to remove excess EtOH and chloroform from the silica membrane, and were
13
14 185 stood to dry at room temperature for 5 min. DNA elution was undertaken using 75 µl of
15
16 186 warmed elution buffer at 54°C (James, et al., 2015).
17
18
19
20 187

21
22
23 188 PCR reactions were undertaken in 25 µl volumes containing approximately 40 ng of DNA
24
25 189 template, estimated in triplicate using a nanodrop 2000 spectrophotometer, 1 × Invitrogen
26
27 190 PCR buffer, 1.5 mM MgCl₂, 1 unit of Invitrogen hot start PlatinumTaq DNA polymerase
28
29 191 (Invitrogen cat No: 10966-018), 1 unit of Qiagen Q-solution, 0.5 µl/ng BSA, 0.2 mM dNTP
30
31 192 mix and 0.2 µM primer mix. Amplification conditions used on a Techne TC-4000 thermal
32
33 193 cycler consisted of an initial denaturation at 94°C for 2 min, followed by 35 cycles of 94°C for
34
35 194 1 min, 55°C for 45 s and 72°C for 1 min finishing with a final extension stage of 72°C for 5
36
37 195 min.
38
39
40
41
42
43 196

44 45 46 197 *Statistical analysis* 47 48

49 198 The probability for exact Hardy-Weinberg proportions, F-statistics and estimates of allele
50
51 199 frequencies between the six sampled subpopulations and each STRUCTURE-identified
52
53 200 cluster were calculated using the program GENEPOP v. 4.2.1 (Rousset & Raymond , 1995;
54
55
56
57
58
59
60

1
2
3 201 Rousset, 2008). Population differentiation between sites was examined using the exact G
4
5 202 test.
6
7
8

9 203

10
11
12 204 Evidence for genetic isolation by distance was assessed by plotting a pairwise genetic
13
14 205 distance matrix (F_{ST}) against a pairwise spatial distance matrix. A Mantel test for dissimilarity
15
16 206 was performed against the two matrices using R v. 3.0.2 (R Development Core Team, 2008)
17
18 207 (permutation = 999 model = strata). Significance values were ascertained using the Monte-
19
20 208 Carlo Markov Chain algorithm (Dememorisation = 1,000, batches = 100, iterations/batch =
21
22 209 1000). Pairwise F_{ST} significance values and Bonferroni p-value corrections for multiple
23
24 210 comparisons were undertaken using the program GENEPOP v. 4.2.1 (Raymond & Rousset,
25
26 211 1995; Rousset, 2008) and R v. 3.0.2 (R Core Team, 2013). The significance of the correlation
27
28 212 between genetic and geographic distances at the individual level was ascertained by Monte-
29
30 213 Carlo simulation (based on 999 replicates) using the R package adegenet v 2.0 (Jombart ,
31
32 214 2008).
33
34
35
36
37
38

39 215

40
41
42 216 The program STRUCTURE v2.3.4 (Pritchard, et al., 2000) was used to estimate rate of
43
44 217 migration and degree of isolation between subpopulations assuming unbalanced and

45
46 218 limited sample sizes (Pritchard et al., 2000). This analysis employs a Bayesian clustering
47
48 219 algorithm to correlate microsatellite allele frequency dissimilarities between individuals
49
50 220 with prior knowledge of sample location. The inclusion of sample location is specifically
51
52 221 recommended when determining low levels of population structuring under small spatial
53
54 222 scales, where a significant F_{ST} value has been determined (Hubisz, et al., 2009). This
55
56
57
58
59
60

1
2
3 223 approach assigns individuals to the most relevant deme based on genetic dissimilarity
4
5 224 between individuals and groups. The admixture model used in this analysis accounts for the
6
7
8 225 possibility of admixture within clusters, as opposed to pure distributions of genotypes, while
9
10 226 remaining robust to the absence of admixture. This method was employed to detect any
11
12
13 227 indication of subtle population structure using the genotype data in this study. The number
14
15 228 of subpopulations, K , was estimated to be between 1 and 6 using a burn-in of 10,000 runs;
16
17 229 Markov Chain Monte Carlo simulation (MCMC) run length of 100,000 with 10 iterations per
18
19
20 230 simulation. Pairwise F_{ST} values between STRUCTURE-identified clusters were calculated
21
22 231 using the program GENEPOP v. 4.2.1 (Rousset & Raymond, 1995; Rousset, 2008) and
23
24 232 examined for significance using the exact G test.

25
26
27 233
28
29
30
31 234 Identification of the number of distinct and genetically consistent groups within the sampled
32
33 235 population was estimated using the rate of change in the log probability of the data
34
35 236 between successive estimates of the number of populations, termed delta K (ΔK) (Evanno,
36
37 237 et al., 2005). The estimation of K was undertaken using the program Structure Harvester
38
39 238 (Earl & von Holdt, 2012). The programs CLUMPP V1.1 (Jakobsson & Rosenberg, 2007) and
40
41 239 DISTRICT v1.1 (Rosenberg, 2004) were then used to produce graphical representations of
42
43
44 240 the structure analysis. However, recent research suggests that unbalanced sample sizes
45
46 241 from known localities may result in the identification of spurious clusters by the program
47
48 242 STRUCTURE (Puechmaille, 2016), which is likely to result in an underestimation of K using
49
50 243 the delta K method outlined in Evanno (2005). As resampling a subset of genotypes to
51
52 244 correct for unbalanced sample sizes is not appropriate in this case due to the small sample
53
54
55 245 size, the approach of identifying a true value for K using the estimators MedMeaK,
56
57
58
59
60

1
2
3 246 MaxMeaK, MedMedK and MaxMedK over 20 repeats per estimation of K was used
4
5 247 (Puechmaille, 2016). The maximum value of K was interpreted by the number of clusters
6
7 248 that contained at least one sampling locality at membership coefficient threshold of 0.5. The
8
9 249 R package Kestimator (Puechmaille, 2016) was used to calculate the estimators listed above.
10
11
12
13

14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45

250
251 We used a cut off assignment to test for the number of potential migrants within each
252 structure-identified cluster (Sacks, et al., 2004). An arbitrary cut off assignment of 70% was
253 selected due to the limited sample size, local spatial arrangement and cluster assignment
254 probability. A χ^2 test was used to assess the difference in the proportion of migrants
255 between clusters.

256

257 The statistical power to reject the null hypothesis of genetic homogeneity in this
258 investigation was assessed by undertaking a power test using the program POWSIM (Ryman
259 & Palm, 2006) at F_{ST} values of 0.001, 0.0025, 0.01, 0.03 and 0.05. Effective population size
260 (N_e), when simulated populations drifted apart, was maintained at 4000 and the number of
261 simulations per run was set to 1000. It is generally regarded that power scores should be
262 greater than 0.8 to be confident of adequate power.

263

264 **Results**

265 *Hardy-Weinberg exact tests and fixation statistics*

266 Genotype frequencies across all loci were found to be in general alignment with Hardy-
 267 Weinberg proportions at the total population level ($\chi^2 = 73.4136$, d.f. = 72, $p = 0.432$). When
 268 examined by locus, three of the 36 tests were shown to deviate significantly from Hardy-
 269 Weinberg proportions across the six sampling localities ($p < 0.05$). However, the exact Hardy-
 270 Weinberg test by population indicated that the majority of this deviation was partitioned to
 271 Site VR1 ($\chi^2 = 33.4919$, d.f. = 12, $p < 0.05$), showing a heterozygote excess at locus cme136
 272 (Weir and Cockerham $F_{IS} = -0.2203$, $p < 0.05$). Weir & Cockerham fixation
 273 statistics indicated that a degree of sub-structuring was apparent in the total population as
 274 highlighted by the multi-locus F_{ST} estimate of 0.0302 (Table 2). Significant genetic
 275 differentiation was apparent between sample sites when examining the variation in allele
 276 frequencies between sites using the exact G test ($\chi^2 = 49.8182$, $df = 12$, $p < 0.05$).

279 **Table 2.** Weir & Cockerham fixation statistics for individual and combined loci across all
 280 localities.

Locus	F_{IS}	F_{ST}	F_{IT}
cme144	0.0819	-0.0080	0.0746
cme136	-0.1783	0.0067	-0.1705
cme206	-0.0024	0.0834	0.0812
cme196	0.0875	-0.0019	0.0858
cme193	0.0223	0.0062	0.0284
cme210	-0.1823	0.1146	-0.0468
All:	-0.0272	0.0302	0.0039

281

282

283

1
2
3 284 *Isolation by distance*

4
5
6 285 Analysis of the entire microsatellite data set found no statistical correlation between
7
8
9 286 Euclidian distance and pairwise F_{ST} values at the population level ($r = -0.1836$, $p = 0.75833$).
10
11 287 Furthermore, no evidence of isolation by distance could be ascertained at the individual
12
13 288 level when the correlation between distance matrices was compared to simulated values
14
15 289 under the absence of spatial autocorrelation (simulated p-value: 0.707, Fig. 2).

16
17
18
19 290

20
21
22 291 *Figure 2 approximately here.*

23
24
25 292

26
27
28 293 *Analysis of population structure*

29
30
31 294 The analysis of genetic variation within and between individuals and sites using the Evanno
32
33 295 method (2005) indicated that the number of ancestral populations genetically represented
34
35 296 in the contemporary data set can be inferred as $K = 3$ (Fig. 3).

36
37
38
39 297

40
41
42 298 *Figure 3 approximately here.*

43
44
45 299

46
47
48 300 STRUCTURE analysis indicated that the population structuring, highlighted by the inbreeding
49
50 301 coefficient ($F_{ST} \approx 0.03$), was largely partitioned between feeding site VR1 and feeding site
51 1

52
53 302 VR2, being consistently dissimilar to each other and the remaining four sites in individual
54
55 303 population assignment. Individuals from the remaining four control sites (C1, C2, C3 and C4)
56
57 304 showed variable population assignment probabilities, thus a high degree of genetic
58
59
60

305

306 admixture was inferred across these sites. The analysis of MedMeaK, MaxMeaK, MedMedK

307 and MaxMedK indicates that the true value of $K = 3$.

308

309 Allelic richness, observed and expected heterozygosity, F_{IS} and the Hardy-Weinberg test for

310 heterozygote excess and the proportion of potential migrants for each STRUCTURE-identified

311 cluster are shown in Table 3. Contrary to our predictions a greater proportion of migrants

312 were found in the STRUCTURE-identified cluster that included the four control sites (Cluster

313 3) when compared with the two supplementary feeding sites ($\chi^2 = 13.091$, $df = 2$, $p < 0.05$).

314

315 **Table 3.** Genetic diversity estimators and proportion of migrants for each STRUCTURE-identified cluster.

Cluster	Site	N	Ar	HO	HE	Overall FIS	HWE (p-value)	Migrants (%)
1	VR1	27	47	103	105.2368	0.0217	0.7322	25.9
2	VR2	19	36	65	64.3377	-0.0103	0.6633	36.8
3	C1,C2,C3,C4	19	43	91	82.8843	-0.1009	0.3053	57.8

316

317 All pairwise F_{ST} values for each STRUCTURE-identified cluster (Table 4) were shown to be

318 significantly different ($p < 0.05$).

319

320 **Table 4.** Pairwise F_{ST} values for each STRUCTURE-identified cluster.

Clusters	Sites	Pairwise F_{ST}
1 + 2	VR1 + VR2	0.0329
1 + 3	VR1 + (C1, C2, C3, C4)	0.0274
2 + 3	VR2 + (C1, C2, C3, C4)	0.0647

1
2
3 321 *Analysis of statistical power*
4
5

6 322 Power analysis undertaken using the program POWSIM indicated that the suite of
7
8 323 microsatellite loci used in this investigation were suitable for differentiating population
9
10 324 structure at F_{ST} values of 0.03 and above, with a Fisher's exact test statistic > 0.8 . Power
11
12 325 analysis with F_{ST} values of 0.001, 0.0025, 0.01, 0.03, and 0.05 were computed as 0.0760,
13
14 326 0.1660, 0.7580, 0.9980 and 1.000 respectively.
15
16
17
18
19 327

20
21
22 328 **Discussion**
23
24
25

26 329 Carnivore spatial organisation is rarely, if ever, homogeneous in space and time. Resource-
27
28 330 based explanations of spatial organisation are able to describe such variation by exploring
29
30 331 the relationship between the availability of resources (e.g. food) and the fitness cost
31
32 332 associated with territorial defence (Johnson, et al., 2001; Johnson, et al., 2002). Theoretical
33
34 333 models that link resource dispersion with spatial organisation describe plasticity in territory
35
36 334 size and stability when the distribution of food is heterogeneous across the environment
37
38 335 (Macdonald, 1983; Johnson, et al., 2002). Thus, traditional explanations of the resource
39
40 336 dispersion hypothesis place emphasis on the selective advantage gained by reducing
41
42 337 territorial defence when the availability of food exceeds the requirements of the individual
43
44 338 and group. For example, studies have concluded that populations of free-ranging red foxes
45
46 339 residing in close proximity to human settlements are more likely to exist at higher densities
47
48
49
50
51

2
3 343

52 340 than their rural counterparts due to the overabundance of anthropogenically derived
53
54 341 sources of food (Bino, et al., 2010). However, the underlying mechanisms by which such
55
56 342 populations are formed and maintained have been heavily veiled by their complexity. In this
57
58
59
60

4
5 344
6
7
8 345
9
10 346
11
12
13 347
14
15 348
16
17 349
18
19
20 350
21
22 351
23
24 352
25
26
27 353
28
29 354
30
31 355
32
33 356
34
35 357
36
37
38
39 358
40
41 359
42
43 360
44
45 361
46
47
48
49 362
50

study we found evidence for a small degree of genetic structuring within the population as a whole (Table 2). Furthermore, a Bayesian analysis of population structure showed that black-backed jackals at supplementary feeding sites were genetically distinct relative to the total population (Fig. 3). However, contrary to our predictions, individuals from the remaining four control sites could not be accurately assigned to a single population of origin based upon their genetic identity alone, and showed a far greater number of potential migrants relative to the supplementary feeding sites (Table 3), which suggests a degree of historic gene flow between these sampling locations. In addition, no evidence of spatial auto-correlation could be detected across the total population (Fig. 2), providing further evidence of a discontinuous population across the sampled area. We believe, therefore, that the results of this study show that far from increasing migration as predicted; clumped, abundant and stable sources of carrion can cause population structuring in the black-backed jackal by reducing gene flow between these sites. However, it should be noted that, while the identification of population sub-structuring is highly indicative of barriers to gene flow within the sampled population, evidence of slight outbreeding (Table 2) suggests that the genetic composition of the total breeding population has not been captured in its entirety. Despite this shortfall, the result of this study provides an informative estimation of the parameters of a population in flux and describes the genetic consequences of a population responding to increased food availability in the resource rich agricultural landscape.

51
52 363
53
54 364
55
56 365
57
58
59
60

Competitive exclusion offers an attractive explanation for the degree of population structuring seen in this study. Once the carrying capacity of the environment has been reached, it is intuitive that a relative increase in competition for food would prompt

1
2
3 366 territorial behaviour and limit or reduce migration and gene flow. Furthermore, due to the
4
5 367 large diversity of alternate sources of prey available to the black-backed jackal within the
6
7 368 agricultural landscape of South Africa (Kamler, et al., 2012), long distance commuting
8
9 369 behaviour, as observed at the CCSR (Jenner, et al., 2001), may not be a cost effective
10
11 370 strategy in this system. Furthermore, investigations into movement patterns of the dingo,
12
13 371 which reside at resource-rich refuse sites in central Australia, have shown that individuals do
14
15 372 not always remain at refuse sites indefinitely. This indicates that further selective pressures
16
17 373 above those predicted by the resource dispersion hypothesis, such as group hunting, may
18
19 374 play an important role in the social structure of the Canidae (Newsome, et al., 2013).
20
21 375 However, given that approximately 24-33% of offspring of territory-holding black-backed
22
23 376 jackals have been recorded as delaying dispersal to provide alloparental care to subsequent
24
25 377 kin (Ferguson, et al., 1983; Moehlman, 1983; Moehlman, 1986; Moehlman, 1987; Estes,
26
27 378 1991), a more likely explanation for the results of this study is that following a substantial
28
29 379 increase in local food availability the offspring of individuals in close proximity to
30
31 380 supplementary feeding sites have reduced dispersal rates, due to the high carrying capacity
32
33 381 of the environment and reduced competition for resources between siblings, resulting in
34
35 382 the formation of genetically distinct clusters of individuals. Previous studies have shown that
36
37 383 pup survival rate is positively correlated to both food availability and alloparental care
38
39 384 (Estes, 1991; Moehlman, 1987). Furthermore, the mechanisms dictating whether an
40
41 385 individual chooses to disperse from its natal range or to remain and act as a helper has been
42
43 386 correlated to food availability, competition for available resources and persecution
44
45
46
47
48
49
50
51
52 387 (Moehlman, 1987; Ferguson, et al., 1983; Minnie, et al., 2016). Therefore offspring that have
53
54 388 failed to disperse from their natal range, in combination with an increase in dispersing
55
56 389 offspring following disturbance from persecution at the control sites (Minnie, et al., 2016),
57
58
59
60

1
2
3 390 would explain, at least in part, the degree of population structuring seen in this study.
4
5 391 However, although previous studies have suggested that a breakdown in territorial stability
6
7 392 can result from clumped and abundant sources of food (Hiscocks & Perrin, 1988; Johnson, et
8
9 393 al., 2002; Bino, et al., 2010), by sampling faeces for genetic material, a prominent territorial
10
11 394 marker in many mammalian species, it is possible that transient individuals may have eluded
12
13 395 genetic identification and potentially induced a sampling artefact to the analysis.
14
15 396 Furthermore, the limited number of microsatellite loci used in this investigation has the
16
17 397 potential to induce a type-two statistical error in this analysis as statistical power is often
18
19 398 reduced when both sample size and microsatellite loci are limited in number. To date, only
20
21 399 six microsatellite markers have been published for the black-black Jackal, which is relatively
22
23 400 few by current standards. However, despite the limited resolution these markers offer for
24
25 401 population structure analysis, they appear to be sufficient for identifying weak
26
27 402 differentiation ($F_{ST}=0.03$), which we regard as still biologically meaningful. It is therefore
28
29 403 recommended that future studies focus on the characterisation of further microsatellite loci
30
31 404 with the aim of undertaking pedigree analysis using high quality tissue samples to accurately
32
33 405 infer relatedness between individuals at supplementary feeding sites.
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48

407 **Conclusions**

1
49 408 Many previous studies have shown that excess food availability can dramatically affect the
50
51 409 population dynamics of carnivores (Hiscocks & Perrin, 1988; Fedriani, et al., 2001; Jenner, et
52
53 410 al., 2001; Johnson, et al., 2001; Bino, et al., 2010; Rotem, et al., 2011; Newsome, et al.,
54
55 411 2013; Yarnell, et al., 2014). An increase in the abundance and density of local
56
57
58
59
60

2
3 412
4
5 413 subpopulations is therefore expected following a substantial increase in carrion availability.
6
7
8 414 The results of this study indicate that anthropogenically provisioned resources (e.g. carrion)
9
10 415 results in genetically identifiable groups of black-backed jackals that show a degree of
11
12 416 historic isolation from the surrounding population. Whether through kin selection or the
13
14 417 principles of competitive exclusion, the formation of a structured population in response to
15
16 418 excess carrion is not unexpected given the assumed territorial breakdown described by the
17
18 419 resource dispersion hypothesis. However the degree of genetic admixture at site VR1
19
20 419 suggests that immigration may play a substantial role in the formation of this cluster. Yet
21
22 420 the ability to identify genetically distinct groups, in response to a vastly increased local
23
24 421 carrying capacity, provides additional insight into the group dynamics of a monogamous and
25
26 422 territorial carnivore in the human-modified landscape.
27
28
29 423
30
31
32
33 424

425 **Acknowledgements**

34
35
36
37 425 We are particularly indebted to the Earthwatch Institution for providing funds to undertake
38
39 426 both field and laboratory work for this study. We are grateful to North West Parks and
40
41 427 Tourism Board, African Explosives Ltd, Standard Bank, Investec Bank, P. Haasbroek, R. Kotze
42
43 428 and K. Burger for written permission regarding land access to collect fresh faecal samples of
44
45

46
47 429 the black-backed jackal in game farm environments. Furthermore, we would like to thank D.
48
49 430 Mactavish, L. Mactavish and J. Martin for the coordination of Earthwatch project “the
50
51 431 scavengers of South Africa”, from which a number of genetic samples were recovered. We
52
53 432 would also like to thank the numerous Earthwatch volunteers that helped identify and
54
55
56
57
58
59
60

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

433 locate faecal samples in the field. DEFRA import permits numbers for genetic material:

434 TARP/11/392, TARP/2012/252 and TARP/12/404.

435

436

437

438

439

440

441

442

443

444

445

446

Review Copy

References

- 447
- 448 Bellan, S.E., Cizauskas, C.A., Miyen, J., Ebersohn, K., Küsters, M., Prager, K.C., Van Vuuren, M.,
449 Sabeta, C. & Getz, W.M. (2012). Black-backed jackal exposure to rabies virus, canine distemper and
450 *Bacillus anthracis* in Etosha National Park, Namibia. *J. Wildl. Dis.* **48**, 371-381.
- 451 Bino, G., Dolev, A., Yosha, D., Guter, A., King, R., Saltz, D. & Kark, S. (2010). Abrupt spatial and
452 numerical responses of overabundant foxes to a reduction in anthropogenic resources. *J. Appl. Ecol.*
453 **47**, 1262-1271.
- 454 Contesse, P., Hegglin, D., Gloor, S., Bontadina, F. & Deplazes, P. (2004). The diet of urban foxes
455 (*Vulpes vulpes*) and the availability of anthropogenic food in the city of Zurich, Switzerland. *Mamm.*
456 *Biol.* **69**, 81-95.
- 457 Earl, D. A. & von Holdt, B. M. (2012). STRUCTURE HARVESTER: a website and program for visualizing
458 STRUCTURE output and implementing the Evanno method. *Conserv. Genet. Resour.* **4**, 359-361.
- 459 Estes, R. (1991). *The behaviour guide to African mammals*. Berkeley: University of California Press.
- 460 Evanno, G., Regnaut, S. & Goudet, J. (2005). Detecting the number of clusters of individuals using the
461 software STRUCTURE: a simulation study. *Mol. Ecol.* **14**, 2611-2620.
- 462 Fedriani, J.M., Fuller, T.K. & Sauvajot, R.M. (2001). Does availability of anthropogenic food enhance
463 densities of omnivorous mammals? An example with coyotes in southern California. *Ecography.* **24**,
464 325-331.
- 465 Ferguson, J.W., Nel, J.A. & De Wet, M.J. (1983). Social organization and movement patterns of Black-
466 backed jackals (*Canis mesomelas*) in South Africa. *J. Zool.* **199**, 48-502.
- 467 Fourie, R.M., Tambling, C.J., Gaylard, A. & Kerley, G.H.I. (2015). Short-term foraging responses of a
468 generalist predator to management-driven resource pulses. *Afr. J. Ecol.* **53**, 521-530.
- 469 Ginsberg, J.R. & Macdonald, D.W. (2004). *Canids: Foxes, Wolves, Jackals and Dogs. Status Survey and*
470 *Conservation Action Plan*. Cambridge: IUCN.
- 471 Hiscocks, K. & Perrin, M. (1988). Home range and movements of black-backed jackals at Cape Cross
472 Seal Reserve, Namibia. *S. Afr. J. Wildl. Res.* **18**, 97-100.
- 473 Hubisz, M.J., Fakush, D., Stephens, M. & Pritchard, J.k., (2009). Inferring weak population structure
474 with the assistance of sample group information. *Mol. Ecol. Resour.* **9**, 1322-1332.
- 475 Humphries, B., Ramesh, T. & Downs, C.T. (2016). Seasonal diet of the black-backed jackal (*Canis*
476 *mesomelas*) on farmlands of KwaZulu-Natal Midlands, South Africa. *Mammalia*.
- 477 Jakobsson, M. & Rosenberg, N.A. (2007). CLUMP: a cluster matching and permutation program for
478 dealing with label switching and multimodality in analysis of population structure. *Bioinformatics.*
479 **23**, 181-1806.

- 1
2
3 480 James, R.S., James, P.L., Scott, D.M. & Overall, A.D.J. (2015). Characterization of six cross-species
4 481 microsatellite markers suitable for estimating the population parameters of the black-backed jackal
5 482 (*Canis mesomelas*) using a non-invasive genetic recovery protocol. *Cogent Biol.* **1**.
- 6
7
8 483 Jenner, N., Goombridge, J. & Funk, S.M. (2001). Commuting, territoriality and variation in group and
9 484 territory size in a black-backed jackal population reliant on a clumped, abundant food resource in
10 485 Namibia. *J. Zool.* **284**, 231-238.
- 11
12 486 Jenner, N. (2008). *Contributions to offspring care by parents and helpers, and the factors affecting*
13 487 *their levels of contribution, in the black-backed jackal Canis mesomelas*. Ph.D. thesis, Zoological
14 488 Society of London's Institute of Zoology, London and University of Kent, Canterbury, UK.
- 15
16
17 489 Johnson, D.D., Kays, R., Blackwell, P.G. & Macdonald, D.W. (2002). Does the resource dispersion
18 490 hypothesis explain group living? *Trends Ecol. Evolut.* **17**, 563-570.
- 19
20 491 Johnson, D.D., Macdonald, D.W., Newman, C. & Morecroft, M.D. (2001). Group size versus territory
21 492 size in group-living badgers: a large-sample field test of the Resource Dispersion Hypothesis. *Oikos*.
22 493 **95**, 265-274.
- 23
24 494 Jombart, T. (2008). Adegnet: A R package for the multivariate analysis of genetic markers.
25 495 *Bioinformatics.* **24**, 1403-1405.
- 26
27
28 496 Kamler, J.F., Klare, U. & Macdonald, D.W., (2012). Seasonal diet and prey selection of black-backed
29 497 jackals on a small-livestock farm in South Africa. *Afr. J. Ecol.* **50**, 299-307.
- 30
31 498 Klare, U., Kamler, J.F., Stenkewitz, U. & Macdonald, D.W. (2010). Diet, Prey Selection, and Predation
32 499 Impact of Black-Backed Jackals in South Africa. *J. Wildl. Manag.* **74**, 1030-1041.
- 33
34 500 Macdonald, D.W. (1983). The ecology of carnivore social behavior. *Nature.* **301**, 379-389.
- 35
36
37 501 Minnie, L., Gaylard, A. & Kerley, G.I. (2016). Compensatory life-history responses of a mesopredator
38 502 may undermine carnivore management efforts. *J. Appl. Ecol.* **53**, 379-387.
- 39
40 503 Minnie, L. (2016). *Effects of lethal management on black-backed jackal population structure and*
41 504 *source-sink dynamics*. Ph.D. thesis, Nelson Mandela Metropolitan University.
- 42
43 505 Moehlman, P.D. (1983). Socioecology of silverbacked and golden jackals (*Canis mesomelas* and *Canis*
44 506 *aureus*). *J. Mammal.* **7**, 423-43.
- 45
46 507 Moehlman, P.D. (1986). *Ecology of cooperation in canids. In: Ecological aspects of social evolution.*
- 47
48 508 Princeton: Princeton University Press, 64-86.
- 49
50 509 Moehlman, P.D. (1987). Social organization in jackals: the complex social system of jackals allows the
51 510 successful rearing of very dependent young. *Am. Sci.* **75**, 366-375.
- 52
53 511 Newsome, T.M., Ballard, G.-A., Dickman, C.R., Fleming, P.J.S. & van de Ven, R. (2013). Home range,
54 512 activity and sociality of a top predator, the dingo: a test of the Resource Dispersion Hypothesis.
55 513 *Ecography.* **36**, 914-925.
- 56
57
58
59
60

- 1
2
3 514 Pritchard, J.K., Stephens, M. & Donnelly, P. (2000). Inference of Population Structure Using
4 515 Multilocus Genotype Data. *Genetics*. **155**, 945-959.
- 5
6 516 Puechmaille, S.J. (2016). The program STRUCTURE does not reliably recover the correct population
7 517 structure when sampling is uneven: subsampling and new estimators alleviate the problem. *Mol.*
8 518 *Ecol. Resour.* **16**, 608-627.
- 9
10
11 519 R Core Team. (2013). R: A language and environment for statistical computing. R Foundation for
12 520 Statistical Computing. Vienna, Austria. URL <http://www.R-project.org/>
- 13
14 521 Ripple, W.J., Estes, J.A., Beschta, L., Wilmers, C.C., Ritchie, E.G., Hebblewhite, M., Berger, J.,
15 522 Elmhagen, B., Letnic, M., Nelson, M.P., Schmitz, J., Smith, W., Wallach, A. & Wirsing, J. (2014). Status
16 523 and Ecological Effects of the World's Largest Carnivores'. *Science*. **343**, 1241484.
- 17
18 524 Ripple, W.J. & Beschta, R.L. (2004), Wolves, elk, willows, and trophic cascades in the upper Gallatin
19 525 Range of Southwestern Montana, USA, *Forest Ecol. Manag.*, **200**, 161-181.
- 20
21
22 526 Rosenberg, N.A. (2004). Distruct 1.1 – Graphical Display of Population Structure. *Mol. Ecol. Notes*. **4**,
23 527 137-138.
- 24
25 528 Rotem, G., Berger, H., Bar (Kutiel), P. & Saltz, D. (2011). The effect of Anthropogenic resources on
26 529 the space-use patterns of Golden Jackals. *J. Wildl. Manage.* **75**, 132-136.
- 27
28 530 Rousset, F. (2008). Genepop'007: a complete reimplementaion of the Genepop software for
29 531 Windows and Linux. *Mol. Ecol. Notes*, **8**, 103-106.
- 30
31
32 532 Rousset, M. & Raymond, M. (1995). GENEPOP (version 1.2): population genetics software for exact
33 533 tests and ecumenicism. *J. Hered.* **86**, 248-249.
- 34
35 534 Rowe-Rowe, D.T. (1983). Black-backed jackal diet in relation to food availability in the Natal
36 535 Drakensberg. *S. Afr. J. Wildl. Res.* **13**, 17-23.
- 37
38 536 Ryman, N. & Palm, S. (2006). POWSIM: a computer program for assessing statistical power when
39 537 testing for genetic differentiation. *Mol. Ecol. Notes*. **6**, 600-602.
- 40
41
42 538 Sacks, B.N., Brown, S.K. & Ernest, B.H. (2004). Population structure of California coyotes corresponds
43 539 to habitat-specific breaks and illuminates species history. *Mol. Ecol.* **13**, 1265-1275.
- 44
45 540 Skinner, J.D. & Chimimba, C.T. (2005). *The mammals of the southern African sub-region*. 3rd ed.
46 541 Cambridge: Cambridge University Press.
- 47
48
49 542 Thorn, M., Green, M., Dalerum, F., Bateman, P.W. & Scott, D.M. (2012). What drives human-
50 543 carnivore conflict in the North West Province of South Africa? *Biol. Cons.* **150**, 23-32.
- 51
52 544 Van der Merwe, I. V., Tambling, C.J., Thorn, M., Scott, D.M., Yarnell, R.W., Green, M., Elissa, C.Z. &
53 545 Bateman, P.W. (2009). An assessment of diet overlap of two mesocarnivores in the North West
54 546 Province, South Africa. *Afr. Zool.* **44**, 288-291.
- 55
56 547 Weir, B. & Cockerham, C. (1984). Estimating F-Statistics for the Analysis of Population
57 548 Structure. *Evolution*, **38**, 1358-1370.

1
2
3 549 Woodroffe, R., Thirgood, S. & Rabinowitz, A. (2005). *People and wildlife, conflict or co-existence*. 9
4 550 ed. Cambridge: Cambridge University Press.

5
6 551 Yarnell, R.W., Phipps, L.W., Dell, S., MacTavish, L.M. & Scott, D.M. (2014). Evidence that vulture
7 552 restaurants increase the local abundance of mammalian carnivores in South Africa. *Afr. J. Ecol.* **53**,
8 553 287-294.

9
10
11 554 Yarnell, R.W., Phipps, L.W., Burgess, L.P., Joseph, E.A., Harrison, S.W.R., Dell, S., MacTavish, D. &
12 555 MacTavish, L.M., Scott, D.M. (2013). The influence of large predators on the feeding ecology of two
13 556 African mesocarnivores: the black-backed jackal and the brown hyaena. *S. Afr. J. Wildl. Res.* **43**, 155-
14 557 166.

15
16
17 558 Zulu, G.C., Sabetta, C.T. & Nel, L.H. (2009). Molecular epidemiology of rabies: Focus on domestic dogs
18 559 (*Canis familiaris*) and black-backed jackals (*Canis mesomelas*) from northern South Africa. *Virus Res.*
19 560 **140**, 71-78.

20
21 561

22
23
24 562

25
26
27
28 563

29
30
31 564

32
33
34
35 565

36
37
38 566

39
40
41
42 567

43
44
45 568

46
47
48
49 569

50
51
52 570

53
54
55 571

56
57
58
59

60

1
2
3 572 **Figures**
4
5
6

7 573 **Figure 1.** *A map depicting the six study sites and the three subpopulations of black-backed*
8
9 574 *jackals sampled in this investigation. Subpopulations are denoted by black circles.*

10
11
12 575 **Figure 2.** *Genetic distance as a function of geographic distance between individual black-*
13
14 576 *backed jackals showing the initial correlation (dot) and the distribution of simulated data*
15
16
17 577 *under the absence of Isolation by distance.*

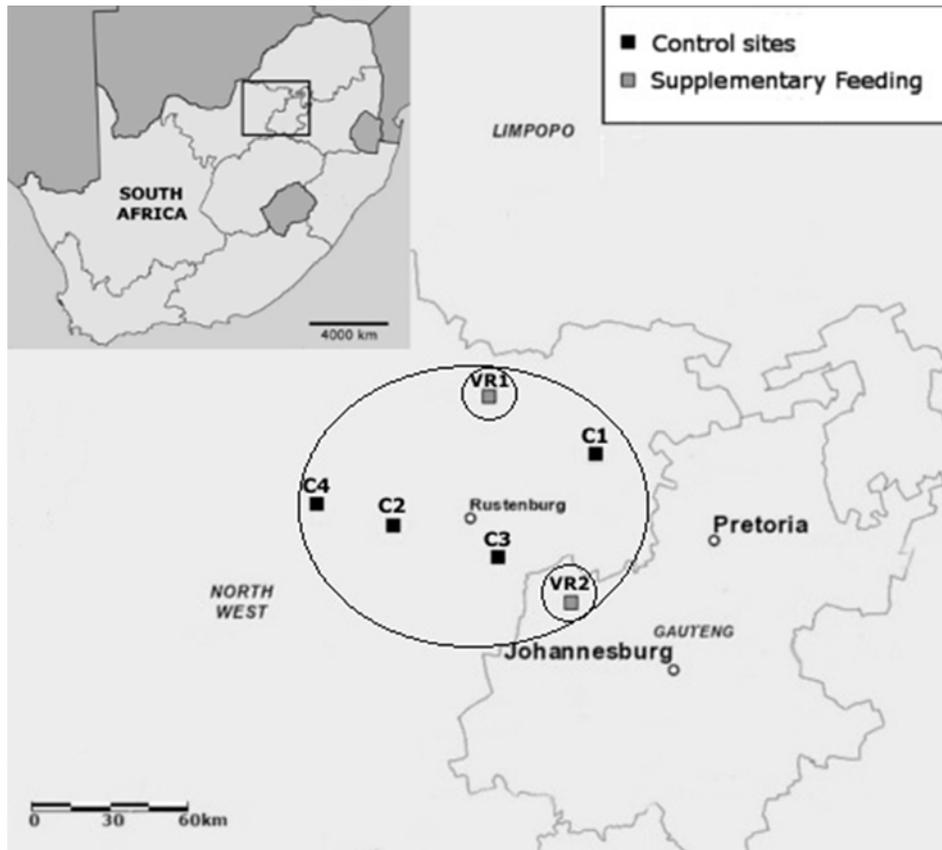
18
19
20 578 **Figure 3.** *A graphical representation of population structure. Individual black-backed jackals*
21
22 579 *are represented by vertical lines, with the population assignment represented in grayscale, k*
23
24
25 580 *= 3.*

26
27
28 581 **Supplementary material A.** *Maps depicting the spatial arrangement of faeces collected for*
29
30 582 *genetic analysis within each game farm site. Faecal deposits of the black-backed jackal are*
31
32 583 *denoted by black circles and carrion stations are represented by white circles.*

33
34
35
36 584 **Supplementary material B.** *Individual black-backed jackal population assignment values for*
37
38 585 *each structure identified cluster.*

39
40
41 586
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59

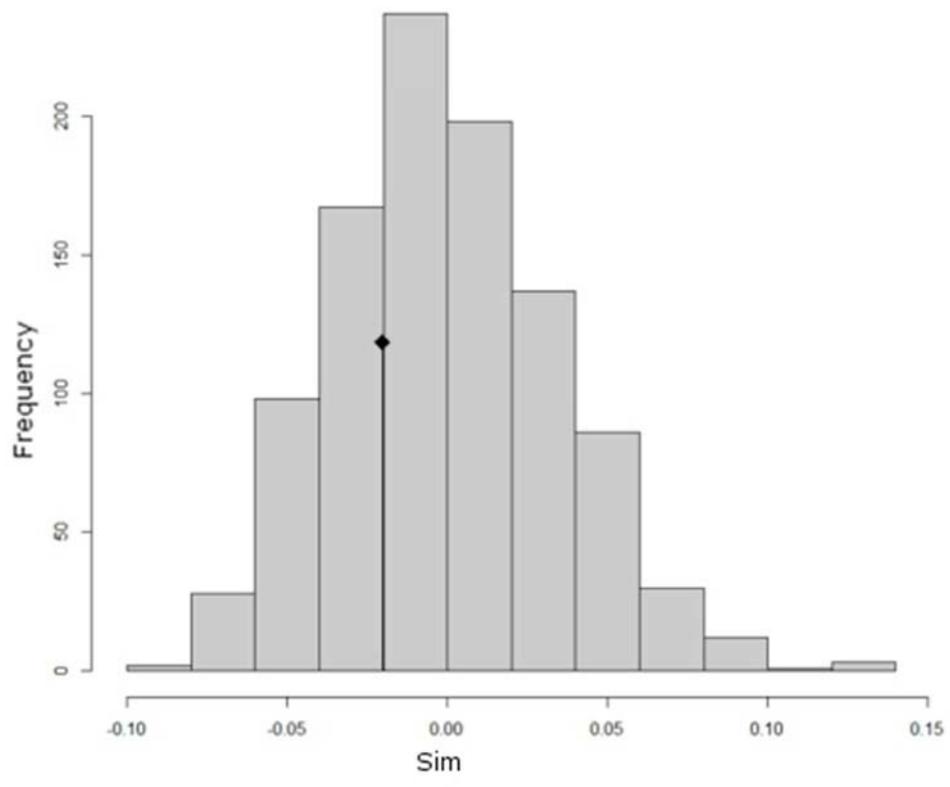
1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60



A map depicting the six study sites and the three subpopulations of black-backed jackals sampled in this investigation. Subpopulations are denoted by black circles.

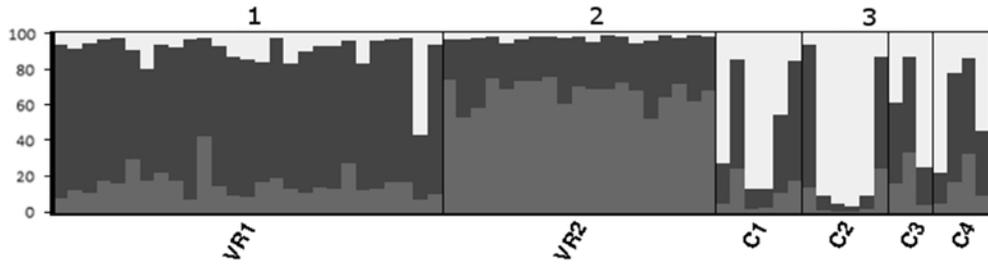
Fig. 1
167x149mm (72 x 72 DPI)

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60



Genetic distance as a function of geographic distance between individual black-backed jackals showing the initial correlation (dot) and the distribution of simulated data under the absence of isolation by distance.
Fig. 2

167x141mm (72 x 72 DPI)



A graphical representation of population structure. Individual black-backed jackals are represented by vertical lines, with the population assignment represented in grayscale, $k = 3$.

Fig. 3

197x54mm (95 x 95 DPI)

Review Copy

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44

45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60



Cover image
Cover image
531x366mm (72 x 282 DPI)

Copy